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Catch rates and distribution pattern of the silky shark, Carcharhinus falciformis, caught by the Taiwanese large-scale longline fishery in the Indian Ocean

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RESEARCH ARTICLE

Catch Rates and Distribution Pattern of the Silky Shark, Carcharhinus falciformis, Caught by the Taiwanese Large-scale Longline Fishery in the Indian Ocean

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Abstract

The silky shark, Carcharhinus falciformis, is widely distributed in tropical and temperate waters, and it is a common bycatch species for tuna longline fisheries. This study examined the distribution of and presents relative abundance indices of the silky shark in the Indian Ocean by using logbook and observer data from the Taiwanese large-scale tuna longline fishery between 2005 and 2019. Due to the high zero catch rate, a zero-inflated negative binomial (ZINB) model was used to standardize catch per unit effort. Due to a lack of detailed targeting information, the fishery strategy was identified by using cluster analysis based on catch composition and then incorporated as an explanatory variable related to the target species in the ZINB model. Size segregation was observed for males and females in the Indian Ocean. Juveniles were mostly concentrated between 10° S and 10° N. Cluster analysis results revealed five fishing clusters based on catch composition that explained the variance in the ZINB models. Our integrated approach improves the understanding of spatiotemporal silky shark dynamics in the Indian Ocean and can be used to derive relative abundance indices for stock assessment and management.

Keywords: Silky shark, CPUE standardization, Indian ocean

1. Introduction

E lasmobranchs (sharks, rays, and skates) are crucial to the marine ecosystem [[1](#page-12-0),[2\]](#page-12-1). These apex predators balance trophic interactions [\[3](#page-12-2)] and sustain the dynamics [\[4](#page-12-3)] of the marine community [\[2](#page-12-1),[5,](#page-12-4)[6](#page-12-5)]. Changes in the abundance of top predators influence the composition of species in the food web [\[1](#page-12-0),[7,](#page-12-6)[8](#page-12-7)]. Numerous studies $[1,9-11]$ $[1,9-11]$ $[1,9-11]$ $[1,9-11]$ $[1,9-11]$ have demonstrated that reductions in the number of sharks and rays lead to a trophic cascade that affects every level of the food chain. Most elasmobranchs are

considered k-selected species characterized by low fecundity, late sexual maturity, slow growth, and long lifespans [\[12](#page-12-9)]. Due to these characteristics, elasmobranchs are more vulnerable to overexploitation than teleost fish are. Moreover, if these species are overfished, their populations require longer recovery times [\[2](#page-12-1),[4,](#page-12-3)[13](#page-12-10)]. According to the results of the International Union for Conservation of Nature (IUCN) Red List assessment [\[14](#page-12-11)], more than 249 elasmobranch species are threatened, and insufficient data existed to classify 487 species. Thus, one-quarter of ray and shark species are classified

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as threatened (critical, endangered, and vulnera $ble-VU$) and data are insufficient for nearly half of the elasmobranch species [\[15](#page-12-12)].

The silky shark, Carcharhinus falciformis, is an oceanic shark with global distribution found between tropical and temperate areas [\[16](#page-12-13),[17\]](#page-12-14). It is a common bycatch species of longline tuna fisheries and the purse seine fishery in the open ocean $[18-20]$ $[18-20]$ $[18-20]$. In many regions, silky shark populations have declined dramatically over the past few decades due to pressures from fishing $[21-23]$ $[21-23]$ $[21-23]$ $[21-23]$ $[21-23]$. Like other elasmobranchs, the silky shark has low fecundity, late sexual maturity, and slow growth, which cause it to be vulnerable to anthropogenic activity and inhibit population recovery [[18\]](#page-12-15). In 2015, the silky shark was listed as VU on the IUCN Red List [[24](#page-12-17)]. In 2016, based on its population, it was listed on the Convention on International Trade in Endangered Species of Wild Fauna and Flora (CITES) Appendix II [[25\]](#page-12-18), indicating that the species is vulnerable. Stock assessments revealed that the silky shark population declined by $46-50\%$ in the Atlantic Ocean between 1992 and 2009 and by 30% in the western Indo-Pacific Ocean between 1995 and 2009 [\[21](#page-12-16),[23\]](#page-12-19). Therefore, retention of the silky shark is banned by International Commission for the Conservation of Atlantic Tunas (ICCAT) and Western and Central Pacific Fisheries Commission. At present, studies on the silky shark in the Indian Ocean are at the preliminary stage, and consequently, the shark's status is uncertain. The stock assessment and management and conservation actions are still inadequate [\[26](#page-12-20)].

Inadequate data are a common problem for shark stock assessment. Due to the low commercial value of the shark, the systemic fishery information needed for modeling is seldom available [\[27](#page-12-21)]. Moreover, data are often undermined by high zero catch rates [[28,](#page-12-22)[29](#page-12-23)]. Because the process of collecting fishery-independent data is often costly and difficult, most Regional Fisheries Management Organizations (RFMOs) rely heavily on catch per unit effort (CPUE) obtained from commercial fishery activity as an indicator of the relative abundance index [\[28\]](#page-12-22). To address this issue, the population trends of bycatch species are commonly estimated by using either delta lognormal models [\[30](#page-12-24)] or by using zero-inflated models [\[29](#page-12-23),[31,](#page-12-25)[32](#page-12-26)]. However, a number of factors influence the CPUE of target or nontarget species including fishing gear, fishing strategies, and fishing operation methods. CPUE standardization is often used to reduce the effects of factors confounding the CPUE index results $[33-36]$ $[33-36]$ $[33-36]$. Lack of data regarding fishing strategies leads to incorrect or biased results. Therefore, cluster analysis based on catch composition is commonly used to detect changes in fishing strategies [[37\]](#page-12-28) and has been widely applied for CPUE standardization by certain RFMOs, such as the Indian Ocean Tuna Commission and ICCAT [\[38](#page-12-29),[39\]](#page-12-30).

Insufficient data exist regarding the age distribution, growth, and reproductive biology of the silky shark in the Indian Ocean [\[40](#page-13-0)]. By using logbook data from Taiwanese vessels operating in the Indian Ocean, we examined the spatiotemporal distribution of the species and calculated its abundance indices. Both of these measures are critical for stock management in the Indian Ocean. This study used a zero-inflated negative binomial (ZINB) model to perform CPUE standardization. The Taiwanese tuna longline fishery data based on catch composition between 2005 and 2019 were clustered to examine fishing strategies and target effects. Observer data were also used to analyze the spatiotemporal distribution of sex and body length. The results derived from this study provide comprehensive information for stock assessment and management of the silky shark in the Indian Ocean.

2. Materials and methods

2.1. Data collection

Logbook and observer data from the Taiwanese large-scale tuna longline fishery (LTLL) between 2005 and 2019 were obtained from the Overseas Fisheries Development Council of the Republic of China. The logbook data of 499,981 longline operations comprising the vessel ID, operation time, operation area, number of hooks, and catches of 18 species including five major tunas, five major billfishes, three sharks, and other species were used to analyze and calculate CPUE ([Table 1\)](#page-4-0). The Taiwanese LTLL fishery operates across the Indian Ocean, and therefore, these fishery statistics are a primary source of information regarding the population status of pelagic sharks. This study also determined biological data for 1591 silky shark individuals obtained by an onboard observer program between 2005 and 2019. The observer data cover an average of 5.67% of all Taiwanese large-scale longline operations in the Indian Ocean and comprise operation time, operation area, and the fork length (FL) of the silky shark. Sex was also determined by examining the external sex organs.

2.2. Spatiotemporal distribution

Catch, effort, nominal CPUE (catch per 1000 hooks), and sex ratio were calculated, and the data

Table 1. Summary of data analyzed for this study by year including the number of sets, total hooks, silky shark catches, and the percentage of silky shark catches among total tuna catches, billfish catches, and observer coverage for Taiwanese large-scale tuna longline vessels in the Indian Ocean from 2005 to 2019.

Year	Set	Hooks	Silky shark catches	observer coverage rate
2005	72,205	229,107,476	7591	0.83%
2006	51,782	165,372,576	2484	1.21%
2007	43,926	140,968,756	2234	5.55%
2008	31,729	102,126,017	3106	5.61%
2009	39,921	128,268,580	4025	5.47%
2010	29,856	97,611,849	1684	7.65%
2011	22,418	72,349,298	929	4.21%
2012	25,206	76,576,911	1935	4.81%
2013	23,719	75,796,412	3069	8.29%
2014	18,475	58,376,963	3098	8.95%
2015	22,535	70,889,449	206	5.99%
2016	31,540	101,456,183	2320	5.34%
2017	29,946	99,221,840	2228	6.32%
2018	28,032	93,060,320	3245	7.53%
2019	28,691	89,907,590	6123	7.25%
Average	33,332	106,739,348	2952	0.06

were then grouped using $5^{\circ} \times 5^{\circ}$ latitude and longitude grids. Hall et al. [[34\]](#page-12-31) suggested that the life history parameters of the silky shark are as follows: neonates $(0-1$ year): 65.98 cm FL; juveniles (age >1 but immature): 76.98-177.51 cm FL for females and 78.35-170.88 cm FL for males; adults (mature, females age >15 and males age >13): >177.51 cm FL for females and >170.88 cm FL for males.

The Indian Ocean was divided into four fishing areas based on the effort distribution and fishing grounds of the target species [[41\]](#page-13-1): (1) Northwest Indian Ocean (north of 10° S, east of 70° E); (2) Northeast Indian Ocean (north of 10° S, 70° E-120 $^{\circ}$ E); (3) Southwest Indian Ocean (south of 10° S, 20° E-60 \degree E); (4) Southeast Indian Ocean (south of 10 \degree S, 60° E-120 $^\circ$ E; [Fig. 1\)](#page-5-0). The annual length-frequency distribution by sex and fishing area was analyzed, and the catches, effort, and nominal CPUE distribution were compared by using both logbook and observer data. Estimates of the sex ratio, life stage, and length-frequency distribution were based on observer data only.

2.3. Cluster analysis

Cluster analysis was based on species composition from logbook data. These species were albacore (ALB), bigeye tuna (BET), yellowfin tuna (YFT), southern bluefin tuna (SBT), billfish, sharks, and others. A two-step method suggested by He et al. [\[37](#page-12-28)] was applied to process the numerous data sets (499,981 sets). The data were aggregated by week and by set to avoid excessive noise caused by clustering operational data. The clusters were then merged with operational set-by-set data by using columns of vessel ID and operation date (year, month, and week) to identify the targeted fishing operations.

For the two-step method, nonhierarchical cluster analysis (K-means method; [\[42](#page-13-2)]) was first applied to group the datasets into 42 clusters based on catch composition ($P_2^7 = 42$; two species can be chosen with priority from seven species). Ward's agglomerative hierarchical cluster analysis was applied to the dissimilarity matrix to calculate the squared Euclidean distances based on the mean species composition from the 42 nonhierarchical clusters. In this study, the clusters were defined as groupings such that the difference in the relative variance between groups and within group was >50% [[43\]](#page-13-3).

2.4. CPUE standardization

The silky shark is a bycatch species of the Taiwanese LTLL fishery. As shown in [Table 2,](#page-5-1) the logbook datasets contain a high proportion of zero catches (95.67% on average), which may lead to bias during CPUE standardization. To prevent such bias, we adopted a ZINB model.

The probability distribution of a ZINB variable Y is given by

(Equation (1) : Binomial model; Equation (2) : Count model—negative binomial, link $=$ logit)

$$
Pr(Y=0) = \omega + (1 - \omega)(1 - k\lambda)^{1/k}
$$
 (1)

$$
\Pr(Y > 0) = (1 - \omega) \frac{\Gamma(y + 1/k)}{\Gamma(y + 1)\Gamma(1/k)} \frac{(k\mu)^y}{(1 + k\lambda)^{y + 1/k}} \tag{2}
$$

where k is the negative binomial dispersion parameter; λ is the mean of the underlying negative binomial distribution; and ω is the probability of an observation being drawn from the constant distribution that always generates zero.

To remove spatiotemporal influences, several factors were considered including 15 fishing years $(2005-2019)$, four calendar quarters, and four fishing areas. Operational variables such as the number of hooks between floats (deep set: \geq 15 hooks; shallow set: <15 hooks) [\[36](#page-12-32)] and vessel size (Vessel: CT5, CT6, CT7) were also considered and incorporated into the cluster results as effects in the CPUE standardization models. All factors were considered to be categorical variables and were evaluated as explanatory variables for ZINBs.

Fig. 1. Area stratification used in this study based on Taiwanese large-scale tuna longline effort distribution and targeted species as recorded by observers (ALB = albacore; YFT = yellowfin tuna; BET = bigeye tuna; SBT = southern bluefin tuna).

A stepwise method was adopted to choose the best-fit model based on the Akaike information criterion (AIC) [\[44](#page-13-4)] and Bayes information criterion (BIC) [[45\]](#page-13-5). A decrease in AIC or BIC indicated a better fit for the ZINB model. The best model was then used in subsequent analysis. Kleiber and

Table 2. Zero catch and immature percentage of silky shark caught by Taiwanese large-scale tuna longline vessels in the Indian Ocean from 2005 to 2019.

Year	Zero catch percentage		Immature percentage		
	Logbook	Observer	Observer data		
	data	data	Female	Male	
2005	94.49	98.31	42.86	33.33	
2006	97.22	87.98	73.08	65.63	
2007	97.59	88.53	57.14	52.80	
2008	97.02	94.55	65.22	46.51	
2009	95.69	95.24	87.80	87.04	
2010	96.97	94.50	92.11	83.13	
2011	97.60	97.51	100.00	84.62	
2012	95.97	96.84	62.50	92.31	
2013	94.67	99.75	91.30	80.00	
2014	93.74	100.00	93.75	92.86	
2015	99.57	99.36	88.89	80.00	
2016	97.52	98.63	96.08	89.80	
2017	95.72	99.79	100.00	84.21	
2018	92.61	98.91	89.29	50.00	
2019	88.68	99.62	94.39	91.46	
Average	95.67%	96.63%	82.29%	74.25%	

Zeileis [[46\]](#page-13-6) proposed using rootograms for model diagnostics to improve the assessment of the fit of a count regression model. We therefore examined our model through a residual analysis using the rootocount regression model. We therefore examined our
model through a residual analysis using the rooto-
gram function in the R software package "countreg" [\[47](#page-13-7)].

The yearly standardized CPUE series was derived by using the adjusted means (i.e., least squared means) of the year effect parameters. The confidence intervals of the standardized CPUE were calculated by using a bootstrap resampling method based on the best model. The number of bootstrap subsamples was determined by the CPUE sample size each year [\(Table 1](#page-4-0)). The 95% confidence intervals for each year were computed by using a biascorrected percentile method with 10,000 replicates [\[48](#page-13-8)]. The statistical analysis and plotting in this study were performed using R 3.6. [[47\]](#page-13-7). Cluster ¹⁴⁶₁. The staustical analysis and plotting in this
study were performed using R 3.6. [47]. Cluster
analysis was computed using the "kmeans" and "analysis was computed using the "kmeans" and "hclust" functions in the R software. ZINB models were implemented by using the "zeroinfl" function of the "pscl" package.

3. Results

3.1. Sex and length distribution

Spatial differences in the catch and CPUE between logbooks and observer records are displayed

Fig. 2. Silky shark catches, effort, and nominal CPUE distribution of Taiwanese large-scale tuna longline vessels recorded by logbook (a) and observer data (b) in the Indian Ocean from 2005 to 2019.

in [Fig. 2](#page-6-0). The logbook data had better spatial coverage than the observer data sets. Logbook data revealed the fishing distribution of Taiwanese LTLL vessels operating between 25° N and 45° S in the Indian Ocean from 2005 to 2019 ([Fig. 2](#page-6-0)). The highest concentration of fishing effort occurred in equatorial areas (20 \degree N-20 \degree S), whereas the highest number of silky shark catches occurred in the Northwest Indian Ocean [\(Fig. 2](#page-6-0)). The nominal CPUE indicated that the silky shark was spatially distributed in the north and southwest Indian Ocean. In the Southwest Indian Ocean, high catch frequency occurred throughout the year.

A total of 1591 silky shark specimens (799 females and 792 males) were recorded by onboard fishery observers [\(Fig. 3](#page-7-0)). FL ranged from 52 to 332 cm; most females were $120-170$ cm and most males were $120-178$ cm ([Fig. 4](#page-7-1)). Average annual immaturity for females and males was 82.29% and 74.25%, respectively ([Table 2\)](#page-5-1). Despite a high proportion of immature individuals, a clear trend in the size distribution of the silky shark was not observed during the study period [\(Fig. 5](#page-8-0)). Additionally, no clear sex segregation was observed [\(Fig. 6](#page-8-1)). The sex ratio deviated from 1:1 between 20° N and 20° S, and a significant difference was observed in sex ratios for

Fig. 3. Distribution of female and male silky sharks by life stage in the Indian Ocean as recorded by observers.

Fig. 4. Size frequency distribution of male and female silky sharks recorded by observers in the Indian Ocean. Vertical lines represent median size at maturity (solid line: female; dashed line: male).

sharks in the south Indian Ocean ([Fig. 6;](#page-8-1) chi-square: 23.9, $df = 3$, $p < 0.001$).

3.2. Cluster analysis

Due to the numerous data sets contained in the logbooks, two-step cluster analysis was necessary to classify the data sets according to target species and fishing strategy. Cluster analysis was used to group the data into five distinct fishing clusters according to the percentage of target species ([Fig. 7\)](#page-8-2). [Table 3](#page-9-0) displays the species composition of each cluster: Cluster 1: Other fishes (OTH); Cluster 2: Yellowfin tuna (YFT); Cluster 3: Bigeye tuna (BET); Cluster 4: Albacore (ALB); Cluster 5: bigeye tuna (BET) and other fishes (OTH). Cluster 2 decreased during the study period, whereas Cluster 4 increased from 2008 onward [\(Fig. 8](#page-9-1)).

3.3. CPUE standardization

ZINB was applied to standardize the CPUE. The best model was selected according to the AIC and BIC. The best-fit model of ZINB was the model with the lowest AIC (234,536) and BIC (235,170) values. This model incorporated all effects. The AIC and BIC values used for model selection are displayed in [Table 4.](#page-9-2) ΔAIC and ΔBIC indicated the reduction in the absolute value of AIC and BIC between the bestfit ZINB model and each other scenario. All variables were statistically significant. The most influential effect was year, followed by area. The smallest effect was observed for quarter. The annual standardized CPUE and nominal CPUE values are displayed with 95% confidence intervals in [Fig. 9](#page-10-0) and [Table 5](#page-10-1). Similar trends were observed: a steady rise between 2011 and 2014, a sharp decrease in 2015, and then a rapid increase from 2016 to 2019. The residual plots indicated that the ZINB models had an excellent fit with the bycatch data [\(Fig. 10](#page-11-0)).

4. Discussion

4.1. Distribution patterns

The silky shark has a variety of habitats and is often captured on the continental shelf and in the open ocean. The shark prefers waters above 23 \degree C [\[17](#page-12-14)]. The majority of silky sharks observed in this study were captured in the north and northwestern Indian Ocean between 20 $\,^{\circ}$ N and 20 $\,^{\circ}$ S and had a high rate of both female and male immature individuals. Previous studies have indicated that the silky shark exhibits some size segregation [\[18](#page-12-15),[49\]](#page-13-9). Newborns and young juveniles are demersal,

Fig. 5. Size distribution of silky sharks in the Indian Ocean by year from 2005 to 2019. Horizontal lines represent median size at maturity (solid lines: female; dashed lines: male).

tending to stay in shelf water nursery areas and deeper parts of the continental and insular shelves, whereas adults are pelagic, moving beyond the continental shelf and returning to shelf waters seasonally to feed and reproduce [\[50](#page-13-10),[51\]](#page-13-11).

A high rate of immaturity was observed across our study area. Except for male samples in 2018, average body size decreased steadily from 2012 to 2019. Hutchinson [\[52](#page-13-12)] indicated that numerous juveniles (<190 cm Total Length) and adults caught by tuna and billfish fisheries were captured at higher latitudes [[11](#page-12-33)[,53](#page-13-13),[54\]](#page-13-14), which is consistent with our findings. However, no clear size segregation was observed in our study. The lack of size segregation may be due to our relatively small sample size from observer data or due to different gear selection or bait types in the areas observed [[55\]](#page-13-15). Data from a longer time series and broader geographical coverage are necessary to understand the reason for this discrepancy.

Larger silky shark specimens of both sexes were found in the southern Indian Ocean, whereas smaller specimens were more frequently observed in tropical and temperate areas. The seasonal movement of the silky shark has been documented. For example, in the Pacific Ocean, Strasburg [\[11](#page-12-33)] demonstrated that silky sharks tend to move from the equator toward higher latitudes in summer. In the Indian Ocean, numerous silky sharks were

Fig. 6. Sex ratio distribution of silky sharks in the Indian Ocean from 2005 to 2019.

Fig. 7. Dendrogram of 42 nonhierarchical clusters for 499,981 longline sets of the Taiwanese large-scale tuna longline fishery in the Indian Ocean from 2005 to 2019.

Species group		Cluster					
				3	4	5	
Albacore (ALB)	Thunnus alalunga	9.71	1.63	8.17	56.87	13.61	
Bigeye tuna (BET)	T. obesus	6.53	20.30	39.88	6.49	19.76	
Yellowfin tuna (YFT)	T. albacares	2.64	56.48	21.24	3.17	10.67	
Southern bluefin tuna (SBT)	T. maccoyii	0.93	0.02	0.37	3.86	1.04	
Swordfish (SWO)	Xiphias gladius	1.46	3.81	4.95	1.52	3.09	
Shark		1.77	2.09	3.83	1.94	3.31	
Others		76.96	15.68	21.57	26.15	48.51	
number of sets		180.016	73,267	145,148	15,561	85,968	
% of Sets		36.01	14.65	29.03	3.11	17.19	

Table 3. Species composition percentage of each cluster from the Taiwanese large-scale tuna longline fishery in the Indian Ocean (2005–2019).

observed in the Gulf of Aden during the late spring and summer [\[56](#page-13-16)]. Additionally, a tagging study of the silky shark indicated sexual segregation and distinct habitat usage among individuals [[55\]](#page-13-15). In our study, a larger number of males were captured in the southwestern Indian Ocean than females were. Neither logbook nor observer data were available for the southeastern Indian Ocean where Bonfil [[56\]](#page-13-16) found silky shark in great abundance. Additional research aiming to refine estimates of abundance and to ascertain movement patterns would

Fig. 8. Annual catch and effort distribution of the five clusters reflecting the targeting strategy of the Taiwanese large-scale tuna longline fleet from 2005 to 2019. Cluster 1: Other fishes (OTH); Cluster 2: Yellowfin tuna (YFT); Cluster 3: Bigeye tuna (BET); Cluster 4: Albacore (ALB); and Cluster 5: Bigeye tuna (BET) and other fishes (OTH).

considerably improve the understanding of silky shark population dynamics in the Indian Ocean.

4.2. CPUE standardization

The logbook and observer data provided valuable spatiotemporal information related to fishing activities. Because logbook data have wider coverage than observer data, the CPUE of the silky shark was standardized based on the logbook data. To ensure that CPUE is proportional to abundance, possible confounding factors must be removed. A variety of methods are available for this task. In a study of oceanic whitetip shark bycatch of the Hawaiian pelagic longline fishery, Brodziak and Walsh [\[57](#page-13-17)] applied five different standardization models: Poisson, negative binomial, zero-inflated Poisson, ZINB, and delta-gamma. The results indicated that a zeroinflated model is more suitable when the zero catch rate of shark is high. Due to insufficient catch process information and a large observed zero catch rate, the ZINB model was deemed appropriate for this study.

Although delta models have also been widely applied for CPUE standardization of nontarget species with high zero catch rates [[36](#page-12-32)], these models were not used in our study because zero-valued observations may be incorrectly pooled [\[29](#page-12-23)]. In this

Table 4. Deviance table for the ZINB model of the silky shark in the Indian Ocean. The absolute value of the AIC and BIC for the null model was 246,223 and 236,190, respectively.

Zero-inflated negative binomial						
Source	Df	Chisg	Pr(> Chisq)	∆BIC	\triangle AIC	
Year	14	689.369	$< 2.2e-16$	-6113.5	-6424.9	***
Ouarter	3	91.594	$< 2.2e - 16$	-98.9	-165.6	***
Area	3	65.137	4.69E-14	-2467.2	-2533.9	***
Cluster	4	132.723	$< 2.2e - 16$	-498	-586.9	***
NHBF	1	103.583	$< 2.2e-16$	-403.3	-425.5	***
Vessel	っ	14.441	0.0007316	-359.4	-403.8	***

Signif. codes: 0 '***', 0.001 '**', 0.01 '*', 0.05 '.', 0.1' ', 1.

Fig. 9. Nominal and standardized CPUEs (per 1000 hooks) with 95% confidence interval for a ZINB model of silky shark abundance.

study, year was observed to have the most significant influence and thus was the most important factor. However, no clear difference was observed between nominal and standardized CPUE. The lack of difference may be due to the small number of years in the data sets and the lack of a homogeneous fishing strategy distribution both spatially and temporally [[58\]](#page-13-18).

Recent studies $[59-62]$ $[59-62]$ $[59-62]$ $[59-62]$ $[59-62]$ have suggested that spatiotemporal models (e.g., the vector autoregressive spatiotemporal model) may yield more precise, biologically reasonable, and interpretable estimates of abundance than conventional methods such as generalized linear models (GLMs) or deltageneralized linear mixed models. Although these models may reduce bias associated with sample selection and fill in spatial gaps associated with fishery-dependent data $[63, 64]$ $[63, 64]$ $[63, 64]$, the model configurations and results of such sophisticated methods are complex and may be difficult to understand [[61\]](#page-13-22). The simple ZINB model adopted in this study generates results that are easily interpreted and understood.

Longline fisheries often adopt different strategies for different target species. Strategic changes include different hook size, gear, operational time, location, and depth. When this detailed information is not available or recorded, cluster analysis is useful to separate the data into different groups based on target species [\[65](#page-13-23)]. Our results indicate that cluster was an important factor explaining the variance of ZINB models. Our integrated approach can be used to understand the fishery strategies for other shark species and to derive relative abundance indices for stock assessment and management.

4.3. Stock status in the Indian Ocean

Little research has been conducted on the CPUE of the silky shark in the Indian Ocean. Two studies in the eastern Indian Ocean (the main operational area of the Indonesian fishery) [[66,](#page-13-24)[67](#page-13-25)] calculated relative abundance indices of the silky shark by using a GLM to estimate the standardized CPUE. Jatmiko [[66\]](#page-13-24) analyzed observer data for the Indonesian longline fleet from 2006 to 2017, and Simeon [\[67](#page-13-25)] conducted research investigating two fishing ports from 2015 to 2016. Simeon's study [\[67](#page-13-25)] indicated that the silky shark population increased between 2015 and 2016. However, the study also demonstrated higher juvenile mortality associated with smaller fishing vessels operating in coastal

Table 5. Estimated nominal and standardized CPUE (per 1000 hooks) of the ZINB for silky shark caught by the Taiwanese large-scale tuna longline fishery in the Indian Ocean.

Year		Original Values		Bias-corrected bootstrap confidence intervals					
	Nominal	Standardized	Lower CI	Upper CI	Mean	STD	CV		
2005	0.11139	0.11825	0.11047	0.11127	0.13452	0.00669	0.04975		
2006	0.05171	0.05043	0.04446	0.05281	0.05441	0.00338	0.06209		
2007	0.05591	0.05440	0.04557	0.06153	0.05585	0.00413	0.07388		
2008	0.10776	0.09849	0.09408	0.09547	0.11814	0.00975	0.08251		
2009	0.11097	0.10731	0.09919	0.11948	0.10495	0.00557	0.05308		
2010	0.06202	0.06489	0.06052	0.06302	0.07479	0.00470	0.06291		
2011	0.04645	0.05465	0.04681	0.05829	0.05949	0.00460	0.07726		
2012	0.08761	0.08878	0.07773	0.09786	0.08973	0.00509	0.05675		
2013	0.14505	0.13510	0.13537	0.14061	0.12172	0.00627	0.05153		
2014	0.18423	0.17042	0.15960	0.20791	0.15746	0.01268	0.08051		
2015	0.00990	0.00915	0.00774	0.01194	0.00820	0.00113	0.13774		
2016	0.08182	0.08056	0.06864	0.09451	0.08038	0.00658	0.08184		
2017	0.08325	0.08367	0.07635	0.09393	0.08297	0.00435	0.05248		
2018	0.12764	0.12909	0.12277	0.14180	0.12653	0.00449	0.03546		
2019	0.23153	0.24614	0.24681	0.25240	0.22976	0.00735	0.03200		

Fig. 10. Rootogram residuals plot for the ZINB model fit to the silky shark logbook data.

areas. These two studies notwithstanding, detailed information is insufficient, and therefore, stock assessments of the silky shark in the Indian Ocean remain uncertain.

By using data from the Taiwanese large-scale longline fishery from 2005 to 2019, this study is the first to investigate the population abundance of the silky shark across the entire Indian Ocean. The catch number was observed to decrease from 2008 to 2012 (lowest value in 2015) but then increase to a maximum in 2019 [\(Fig. 9](#page-10-0)). In the southwestern Indian Ocean, high catches of silky sharks were frequent throughout the year, with silky sharks caught as bycatch by LTLL targeting oilfish (Ruvettus pretiosus) and escolar (Lepidocybium flavobrunneum). Because catches of both silky shark and other species (primarily oilfish and escolars) were lower in 2015 than in other years, the low catches of silky sharks in 2015 may be due to particularly low fishing effort in the southwestern Indian Ocean. However, these results may reflect only partial stock status because our spatial coverage may be insufficient to judge the entire stock status. Additionally, only commercial fishing records were used to calculate relative abundance indices in this study. Although the indices were derived by using standardized procedures, fishery-independent data, such as survey data, would be more accurate because there would be no catch bias due to discard, release, or nonreporting, which are typical occurrences for bycatch species such as sharks. Further fishery-independent studies are necessary to better evaluate the status, ecology, and distribution of the silky shark in the Indian Ocean. For future management of the silky shark, Tsai [\[27](#page-12-21)] suggested that sex-specific and immature shark protection strategies are the most efficient conservation method. Because global shark catches and landings are increasing, the monitoring of silky shark populations is necessary to ensure the protection of this species in the Indian Ocean.

Conflicts of interest

The named authors have no conflict of interest, financial or otherwise.

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